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## ARTICLE

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# The unequal burden of human-wildlife conflict

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Human-wildlife conflict is one of the most pressing sustainable development challenges globally. This is particularly the case where ecologically and economically important wildlife impact the livelihoods of humans. Large carnivores are one such group and their cooccurrence with low-income rural communities often results in real or perceived livestock losses that place increased costs on already impoverished households. Here we show the disparities associated with the vulnerability to conflict arising from large carnivores on cattle (Bos taurus) globally. Across the distribution of 18 large carnivores, we find that the economic vulnerability to predation losses (as measured by impacts to annual per capita income) is between two and eight times higher for households in transitioning and developing economies when compared to developed ones. This potential burden is exacerbated further in developing economies because cattle keepers in these areas produce on average 31% less cattle meat per animal than in developed economies. In the lowest-income areas, our estimates suggest that the loss of a single cow or bull equates to nearly a year and a half of lost calories consumed by a child. Finally, our results show that 82% of carnivore range falls outside protected areas, and five threatened carnivores have over one third of their range located in the most economically sensitive conflict areas. This unequal burden of humancarnivore conflict sheds light on the importance of grappling with multiple and conflicting sustainable development goals: protecting life on land and eliminating poverty and hunger.

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ne of Earth's foremost challenges to sustainable development is ensuring that the human communities living alongside wildlife and protected areas prosper<sup>1-3</sup>. While wildlife such as large carnivores-species ranging from wolves and hyenas to lions and bears-are critical for regulating ecosystem processes that improve human well-being<sup>4-7</sup> and contribute substantially to tourism economies<sup>8,9</sup>, their protection is challenging because they often kill valuable livestock in both community lands and areas adjacent to protected areas 10-13. The economic shocks of losing livestock to large carnivores can be very high<sup>14</sup>, because as much as two-thirds of a household's annual income can be lost in a single livestock predation event<sup>15</sup>. For example, families living inside Jigme Sigmye National Park in central Bhutan lost on average 17% of their yearly per-capita income from tiger (Panthera tigris) and leopard (Panthera pardus) predation<sup>16</sup>, and those on the edge of Tanzania's Serengeti National Park lost over 19% annually from leopard and lion (Panthera leo) predation<sup>17</sup>. Moreover, some 750 million to one billion livestock keepers own limited livestock, are landless, live on less than two US dollars per day<sup>18,19</sup>, and have pastureland overlapping some of the least productive vegetation zones<sup>20</sup>. In lower latitudes, livestock are also kept as a risk management asset<sup>21</sup>. The financial shocks from human-wildlife conflict often disproportionately affect such producers<sup>15</sup> and are an additional source of stress in communities already impacted by climate change, armed conflict, and disease events<sup>22-24</sup>.

In some areas carnivores are valued for cultural and religious reasons<sup>25-27</sup> but where these are absent conflict between agricultural communities and carnivores often escalates carnivore extinction risk due to the retaliatory killings following livestock predation<sup>28-31</sup>. For example, intense conflicts between cattle farmers and African lions in WAZA National Park in Cameroon and Queen Elizabeth National Park in Uganda have led to unsustainable rates of retaliatory killings over the past three decades, causing severe declines in these lion populations<sup>32,33</sup>. Such retaliatory killings are important for a series of reasons including that they negate the increasingly important role carnivores play in both trophic regulation<sup>34,35</sup>, and the direct benefits they provide to humans<sup>5</sup>. For example, high mule deer (Odocoileus hemionus) densities and reduced tree recruitment in the Zion and Yosemite National Parks USA were linked to localized extinctions of mountain lions (Puma concolor)<sup>36,37</sup>. Similarly, mountain lions and gray wolves (Canis lupus) make roadways for humans safer in South Dakota and Wisconsin USA through reducing vehicle collisions with deer and save millions of dollars in resulting insurance costs and hospital fees<sup>7,38</sup>. Given that rangelands make up over half of the earth's terrestrial surface<sup>39</sup> and that agricultural lands are projected to increase in extent by 2-10 million km<sup>2</sup> (~15-20%) in the coming decades<sup>40</sup>, large carnivores are at unprecedented risk of extinction along with the unique ecosystem services they provide.

While research on the negative economic impacts of large carnivore predation on livestock has been ongoing at local scales<sup>41,42</sup> including some well-developed examples<sup>43,44</sup>, and even those examining spatial human-wildlife conflict risk at the continental scale<sup>45</sup>, no work to date has comparatively explored the potential economic burden of large carnivores globally. Here we present the first spatially explicit analysis of the potential economic burden (Supplementary Box 1) arising from human-wildlife conflict at the global scale illustrating the financial and social costs of losing livestock on human communities. We do this by examining the proportion of GDP per capita (hereafter referred to as per capita income) that is vulnerable from the loss of a single cattle calf (the equivalent of 250 kg or one tropical livestock unit<sup>46</sup>) in 133 countries that overlap with the habitat of 18 large carnivore species known to prey on cattle (adjusted for

the presence of cattle). We then assess the direct and opportunity costs of lost calories to households from the predation of a single cow or bull and contrast this against the production yield of meat per animal. Finally, we examine the proportion of carnivore range found within highly economically sensitive conflict areas (i.e., areas where communities would experience  $\geq 25\%$  loss to per capita income during a predation event, we define these areas as frontline communities) and discuss this in the context of a suite of already highly threatened carnivore species. Our study shows the predicament facing economically fragile households and carnivore species occupying the same habitats, thereby illustrating the difficult tradeoffs between three closely interlinked United Nations Sustainable Development Goals: no poverty (Goal 1), zero hunger (Goal 2), and protecting life on land (Goal 15).

### Results

Potential burden hotspots. Using an analysis based on subnational 1st level administrative regions our data show that the world's poorest people may bear the highest cost of living with large carnivores (Fig. 1). This is based on the disparities in economic vulnerability to carnivore predation on cattle. People living in developing countries will on average experience an eightfold higher potential economic burden ( $\bar{x} = 32\%$ , range = 0.02–201% of per capita income lost in a large carnivore predation event) than those living in developed economies ( $\bar{x} = 4\%$  income lost, range = 1-9%, Supplementary Table 2). Moreover, 13% of the countries (n = 17 out of 133) in carnivore range globally are potentially under threat of losing more than half of their per capita income. The most economically sensitive countries (i.e., those with the highest potential economic burden where >50% economic income loss would occur through predation) in our global assessment were the African states of Mozambique, Guinea-Bissau, Malawi, Tanzania, Uganda, and Burundi, and the Southeast Asian countries of Cambodia and Laos. In these countries human communities are vulnerable to losing all or double their annual income  $(\bar{x} = 132\%)$ , more than range = 106-201% of per capita income) if a single calf is killed by carnivores. The situation is different in some of the most economically developed countries like Sweden, Canada, United States, Australia, and Spain where the vulnerability of a single predation event is nominal ( $\bar{x} = 1.67\%$ , range = 1.29–1.92% of per capita income, Supplementary Table 2).

Carnivore species in highly sensitive conflict areas. The geographic distribution of the 18 large carnivores in our analysis was 72,313,995 km<sup>2</sup> in 2009, and 82% of this fell outside of protected areas. Nearly a quarter of all carnivore range in our analysis (23%) fell within the highest conflict burden areas ( $\geq$ 25% per capita income vulnerable to a single predation event). Ten species had more than a third of their range in areas where a predation event would represent a severe economic burden (Table 1), and this included eight species considered globally threatened by the IUCN. These species had a mean proportion of 61% of their range overlapping areas where a cattle predation would amount to  $\geq 25\%$  of per capita income lost. The species with the highest range overlap were the Snow leopard Panthera uncia (89% range overlap, Fig. 1), African lion Panthera leo (78% range overlap), Asiatic black bear Ursus thibetanus (70%), striped hyena Hyaena hyaena (66%), and leopard Panthera pardus (64%, Table 1). The species with the lowest range overlap with highly sensitive conflict areas were the dingo Canis lupus dingo (0% range overlap), red wolf Canis rufus (0%), American black bear Ursus americanus (0%), puma Puma concolor (1%), and jaguar Panthera onca (4%, Table 1).



**Fig. 1 The average annual per capita income percentage loss recorded across the range of 18 large carnivores globally under a single calf predation event.** Our model assumes that a single cattle keeper is exposed to a single calf (250 kg) predation event anywhere across a given carnivore's range. The orange silhouettes represent those species that have >25% of their range located in areas where communities would experience >50% economic income loss through predation. Bar charts are provided for the ten species of large carnivore identified in our analysis that have more than a third of their range in areas where a conflict event would represent a high economic burden (i.e.,  $\geq$ 25% of per capita income vulnerable from a single predation event). Silhouettes obtained from www.phylopic.org and are used under the Creative Commons Attribution-Non Commercial 3.0 Unported license. Silhouette credits are as follows: *Puma concolor* = Cristian Osorio & Paula Carrera, *Ursus americanus, Ursus thibetanus* and *Ursus arctos* = Tracy Heath, *Canis lupus, Panthera pardus, Acinonyx jubatus, Hyaena* hyaena, *Parahyaena* brunnea, and *Panthera leo* = Margot Michaud, *Panthera* onca, *Panthera uncia* and *Lycaon pictus* = Gabriela Pamono-Munoz, *Canis lupus dingo* = Sam Fraser-Smith, *Panthera tigris* = Steven Traver, *Cuon alpinus* = Michael Keesey, *Canis rufus* = David Orr, *Crocuta crocuta* = Oscar sanisidro.

Food security impacts. Predation events also represent a loss in direct and opportunity costs to cattle keepers. Assuming a single predation of a cow or bull, this can result in 227.33-1229.07 immediate usable kilocalories (1kcal = 1000 calories) lost to a cattle keeper and their family (Fig. 2). This caloric loss is highest in the developed world ( $\bar{x} = 817.80$  kcal, range = 478.17-1130.22) but most severe in developing economies because daily per capita meat intake is 37% lower than in the developed world (Supplementary Table 4). The immediate caloric losses from a single predation event in developing economies translates to nearly 1.5 years ( $\bar{x} = 1.42$ , range = 0.77–3.12) of lost daily calories required by a 2-3-year-old child, and roughly two thirds of a year  $(\bar{x} = 0.69, \text{ range} = 0.26-1.41)$  for a 12-13-year-old adolescent, and 30–59-year-old adult human ( $\bar{x} = 0.66$ , range=0.25–1.35). We estimate another 131.04 and 33.88 kcal are lost annually from potential milk and meat production respectively under the assumption that 312 kg of milk and 12.5 kg of meat are produced annually<sup>47</sup>.

**Economic disparities in national cattle production**. We found that cattle prices globally were comparable for the 2009 year ( $F_{(2,76)} = 0.41$ , p = 0.66, Fig. 3), the same year as our AOH habitat carnivore data. There was also no significant difference in cattle price between developing, transition, and developed economies for multiple years beginning in 1961 ( $F_{(2,103)} = 0.32$ , p = 0.73, Fig. 3). However, we observed disparities in the meat yield per cattle between these different economic groups with developing

(174.05 kg/animal, SD = 63.39) and transition (153.4 kg/animal, SD = 36.99) economies producing significantly less meat per animal ( $F_{(2,131)} = 20.75$ , p = <0.00001) when compared to developed states (253.98 kilograms/animal, SD = 54.86) for the 2009 year, and over multiple years with developing (155.91 kilograms/animal, SD = 53.97) and transition (149.87 kg/animal, SD = 37.11) economies producing significantly less meat per animal ( $F_{(2,6983)} = 827.088$ , p = 0.00) when compared to developed states (223.94 kg/animal, SD = 57.88).

#### Discussion

Our results highlight two key findings, namely that Earth's poorest communities pay the highest financial and food security costs for conflict (it is eight orders of magnitude worse for a cattle keeper, as a proportion of their per capita income, to lose a calf from conflict in a developing country than a developed one), and more than half of the world's largest carnivore species have over one third of their habitat overlapping the most economically vulnerable human communities (what we term economic frontline communities). Our results confirm that not only is this true in absolute per capita income loss, but also due to cattle keepers in developing economies producing less meat per animal and having a lower meat intake per capita. This is notable because in many areas of the developing world cattle keepers are already under immense pressure from localized rainfall patterns, drought, and climate change<sup>21,48</sup>, and the fragmentation of once contiguous pasture lands<sup>49</sup>. This is causing livestock units per capita

Table 1 The proportion of Area of Habitat (AOH) of 18 large carnivore species within low, intermediate, high, and severe vulnerability related to a calf depredation event. This area represents the percentage of each species range that falls within the most economically sensitive communities on earth.

Scientific name	IUCN status	Mean annual income loss (%)	% Range in low- vulnerability area	% Range in intermediate vulnerability area	% Range in high- vulnerability area	% Range in severe vulnerability area
Panthera uncia	VU	41.68	4.59	1.75	5.13	88.52
Panthera leo	VU	59.07	2.78	5.16	13.76	78.29
Ursus thibetanus	VU	31.26	10.09	2.94	17.18	69.79
Crocuta crocuta	LC	47.57	8.38	5.40	18.07	68.15
Hyaena hyaena	NT	36.64	3.67	3.17	27.60	65.57
Panthera pardus	VU	46.81	8.59	8.50	18.56	64.35
Cuon alpinus	EN	33.01	4.18	3.35	30.53	61.93
Acinonyx jubatus	VU	32.51	4.63	15.79	34.54	45.04
Panthera tigris	EN	33.57	20.09	9.34	25.74	44.83
Lycaon pictus	EN	44.79	15.32	8.28	40.55	35.86
Ursus arctos	LC	11.63	24.70	18.66	48.22	8.42
Canis lupus	LC	10.02	33.48	17.92	40.73	7.86
Parahyaena brunnea	NT	12.27	18.93	55.13	22.03	3.91
Panthera onca	NT	13.50	12.18	15.70	68.57	3.55
Puma concolor	LC	8.60	44.33	14.37	39.89	1.41
Canis lupus dingo	LC	1.90	100	0.00	0.00	0.00
Canis rufus	CE	100	0.00	0.00	0.00	0.00
Ursus americanus	LC	1.54	99.14	0.86	0.00	0.00

Classifications are as follows – low vulnerability = 0-5% per capita income loss, intermediate vulnerability = 5-10% per capita income loss, high vulnerability = 10-25% per capita income loss, and severe vulnerability = 225% per capita income loss.



**Fig. 2 The direct and opportunity costs of losing a single cow or bull in areas of low per capita income. A** The direct costs of losing a single cow or bull, and Bthe opportunity costs of losing a single cow or bull. Costs are calculated in direct caloric. Data and references are taken from 47 and 109. Lost opportunity cost may also be manifested as a negative externality whereby a lost cow or bull decreases herd size, reproductive potential, and places livestock producers where they can no longer maintain a pastoral lifestyle as noted in 46 and 48. Silhouettes obtained from www.phylopic.org and www.freepik.com. Silhouette credits are as follows: bottle and meat = macro\_vector on Freepik, human silhouttes = www.publicdomainvectors.com, cattle = Steven Traver and Andreas Preuss.



12000

Cattle prices by economy (all years combined)

7000

Cattle prices by economy (2009 only)

Fig. 3 Box plots representing variation in cattle price variation and cattle meat yield per animal according to either developing, transition, or **developed economies.** We provide data for both the 2009 year (which corresponds to our AOH carnivore habitat data), and for all years since 1961. Silhouettes obtained in public domain.

to decline<sup>50</sup>. This in turn has a knock-on effect on herd size that when below a critical threshold leads pastoralism to no longer be viable<sup>46,48</sup>. Estimates vary but typically if a cattle keeper owns two or fewer adult cattle (4.5 tropical livestock units equivalent to 1125 kg of livestock biomass per capita) they are unable to maintain a nomadic cattle keeping lifestyle<sup>48</sup>. Any source of livestock loss at this point is too severe for economic recovery<sup>46</sup> and will potentially erase the wealth of an entire cattle keeping family. These income losses may have a knock-on effect on food security, and this may disproportionately affect women who tend to forgo meals to keep the household fed<sup>51,52</sup>. This mirrors crop raiding scenarios like those highlighted in the Kavango-Zambezi Transfrontier Conservation Area<sup>53</sup>. These authors showed that elephant-associated crop raiding exacerbated the likelihood of food insecurity due to an already reduced rainfall environment. We anticipate that such knock-on effects we discuss here are likely experienced by many if not most of the households that would lose  $\geq 25\%$  of per capita income in our analysis. Moreover, further knock-on effects are likely, such as migration, subsequent social, and political instability, and impacts to childhood cognitive skills and education due to loss of calories<sup>54</sup>. Our analysis also illustrates two sources of additional and lost opportunity cost, that of useable calories from meat, and those garnered from milk and associated products. We estimate the immediate useable calories to be somewhere between roughly half a year and 1.5 years of calories depending on human age, while we estimate an additional ~300 kg of milk and 12 kg of meat/year from cattle yield. These kinds of costs ignore additional valuations of other cattle products such as dung used for fertilizer<sup>55,56</sup>, the hidden mental costs associated with the loss of a sacred animal in some regions<sup>57</sup>, and critical thresholds of cattle herd size<sup>48</sup>.

Charismatic carnivore species like African lions and tigers are highly valuable to societies, and this is especially true at the national or global level<sup>58,59</sup>, but at more local scales they often have little or even negative value<sup>9,15</sup>. Despite the lamentations of western governments and celebrities over declining carnivore populations, the global community suffers little to no personal cost from large carnivores. This lies in stark contrast to the experiences of local communities living alongside these predators in the developing world for whom a single predation event represents on average a third of their annual income. It is these local communities who "are burdened disproportionately with global wildlife maintenance despite the commonplace notion of wildlife resources as the birthright of people everywhere"<sup>60</sup> (pp. 35-36). Furthermore, outside actors from developed nations at times impede conservation efforts on reserve edges and community lands in developing nations, such as in the case of trophy hunting bans that deny communities a proven mechanism to benefit economically from wildlife<sup>61,62</sup>.

Our results arrive at a critical juncture, where calls for environmental conservation are increasingly tempered by recognition of the environmental justice concerns that go with such requests<sup>60,63,64</sup>. The poorest 40% of humanity suffer impacts from climate change 70% greater than the global average<sup>65,66</sup>. Likewise, our analysis shows that human communities lying outside of protected areas overlap with 82% of the Earth's large carnivore range, whilst also bearing potentially the heaviest economic burdens of living alongside them. Building on work highlighting the tradeoffs between biodiversity conservation and food production<sup>63</sup>, we highlight the at times conflicting nature of three of the United Nations Sustainable Development Goals (SDGs): no poverty (Goal 1), zero hunger (Goal 2), and protecting life on land (Goal 15). In the context of human-wildlife conflict these three SDGs are often in direct competition with one another, mainly because there are no sufficient bridging mechanisms for communities to access the potential value of

damage-causing but charismatic species. In cases where human communities living alongside protected areas and damagecausing wildlife realize the monetary value from these species (e.g., through ecotourism revenues, conservancies, and population performance payments), and this revenue is not subjected to elite capture or is marginal in nature<sup>67,68</sup>, pressures on protected areas and wildlife wane. This is due to increases in human development indices and people escaping from poverty traps<sup>69–71</sup>. Indeed, human-wildlife conflict has also been shown to reduce when non-economic conflict reduction models are used. These include *inter alia* the introduction of human and canine guardians for livestock<sup>41</sup>, and the building of protective enclosures and bomas to shield livestock against predators<sup>72</sup>.

Our results illustrate the urgent need to create economic bridging measures that will reconcile the high value placed on carnivores globally or nationally, with the economic inequalities suffered by local communities. Recent developments in the novel conservation financing sector are an example of such mechanisms. Two examples include South Africa's "Rhino Bond" and Sweden's payment for wolverine presence. Both are tied to the population performance of these species in community lands and protected areas<sup>73,74</sup>. Conservancy models which promote tourism and revenue generation in non-protected areas but are not necessarily limited to non-consumptive-tourism only (still having a livestock component<sup>75</sup>) are another financial model for the problem we have identified here<sup>76</sup>. These solutions deal explicitly with the problem of damage-causing carnivore species (seven of which that are considered globally threatened by the IUCN Red list of threatened species) ranging on non-protected community land which is not only suitable habitat but is also co-inhabited by cattle. It should also be noted that the costs associated with predation events often pale when compared to climatic or disease-related shocks which run into the billions of US\$ dollars<sup>77</sup>.

Our analysis of the potential economic burden arising from losing livestock to carnivores was only possible due to recent developments in global socioeconomic<sup>78</sup> and carnivore range spatial data<sup>79</sup>. We feel however, that our analysis is extremely conservative for several reasons, namely we only show the potential economic ramifications of a single predation event. Carnivore depredation often manifests itself in specific areas due to habitat or ecological variables<sup>80,81</sup>, households often experience predation events multiple times per year<sup>15</sup>, and sometimes carnivores engage in surplus killing<sup>82-84</sup>. Our analysis does not calculate actual depredation rates, it only illustrates highly vulnerable (and buffered) economic areas globally that would suffer under a predation event. Our valuation of a predated calf is also likely low, because we adjusted the measure of FAO market cattle value to the slaughter weight of a single ~6-month-old calf (250 kg), equating to roughly 33% of the economic value of an adult cow, and our measure of a single calf as a proportion of per capita GDP is well below the threshold of total income generated from many cattle production systems in Africa and Asia (for instance across much of sub-Saharan Africa livestock typically contributes between half and all of household income in rural settings, with cattle contributing 70-90%; 47). We acknowledge that a lack of regional cattle price data means that our analysis of cattle market price is relatively coarse. Prices in livestock also change with the breed of cattle, seasonality (and drought) and with prevailing macro and micro economic conditions<sup>85</sup>. Further, our data originate from 2009 and the acceleration of humaninduced climate change, the COVID-19 pandemic, and increased rates of political strife may have impacted the severity of our analysis even further.

How cattle keepers view and utilize cattle also varies sharply across the globe, and as<sup>86</sup> note in their key paper "there is an

enormous variability in herd management strategies, in social organization, in land tenure, degree of dependence on agricultural products, interactions with outside groups, differentiations of sex and age, etc". We agree that such variation may be present even at incredibly fine spatial scales. For example, the Bahima and Karamajong cattle keepers of Uganda's Mbarara and Karamoja districts, place immense cultural and monetary value into cattle keeping. It is a central part of their identity<sup>87</sup>. They largely subsist off cow milk and blood, use cows as dowry or enzhugano<sup>88</sup>, but rarely slaughter the cows for meat<sup>89</sup>. Contrastingly, South African Zulus slaughter a high quantity of cattle for consumption in their diet<sup>90</sup>. Similar contrasts can be observed in India. Roughly 80% of the human population is comprised of Hindus<sup>91</sup> which consider cattle as sacred and do not slaughter and consume cattle. This contrasts with the roughly 14% of Indians which are Muslim. Forty percent of these Muslims include cattle meat in their diet. Our analysis does not capture the nuances of these cultural or even religious valuations of cattle. We also do not differentiate the variations in production practice of cattle globally. For example, in the Brazilian Pantanal cattle ranches are large (over half are between 5 and 30,000 hectares in size<sup>92</sup>, intensive in nature, and slaughter and export to the international market<sup>93</sup>. This contrasts with for example Maasai group ranches in southeast Kenya that own on average between 250 and 650 hectares depending on their location<sup>94</sup>. Our analysis encapsulates both highly productive intensive systems, and nomadic, subsistence production types.

Finally, our results only touch on the many potential opportunity costs (Fig. 3) stemming from conflict. There are a multitude of hidden costs<sup>95</sup> and perceived conflicts that can be associated with such a loss including increased workload to make up for financial losses<sup>51</sup>, physical displacement of households<sup>96</sup>, the physical and disease exposure risks of guarding livestock at night to prevent further losses<sup>95</sup>, transaction costs of pursuing compensation payments, and the failure to obtain fair livestock value<sup>97,98</sup>. A recent example from Zimbabwe shows that the presence of an African lion equates to negative USD\$180 per person per year due to fear, and a lack of trust in compensation by authorities responsible in mitigating conflict events<sup>9</sup>. There may also be a host of psychological effects stemming from such conflict including fear of attack by carnivores, hesitancy to move in the dark, grief over lost livestock, and even PTSD from livestock loss, not to mention the immense cost of human life itself<sup>99</sup>.

### Methods

**Mapping large carnivore habitat**. We mapped the spatial habitat extent of 18 large carnivores known to prey on cattle (Supplementary Table 1) using Area of Habitat (AOH) data from<sup>100</sup>, except for the dingo *Canis lupus dingo*, which was mapped following<sup>101</sup>. AOH represents areas of fine scale (300 m) high habitat suitability within large carnivore IUCN geographic distributions during the year 2009. AOH has been used in several recent global studies<sup>102,103</sup> due to its reduced risk for commission errors<sup>79</sup>, fine spatial scale, and incorporation of heterogeneous environmental variables (e.g., land cover, elevation, and hydrological features<sup>104</sup>. These layers were used for subsequent analyses. Excluded countries are presented in Supplementary Table 3.

**Mapping per capita income**. We mapped annual per capita income at the subnational scale (i.e., state or province scale) from for the year 2009<sup>78</sup>. We chose the year 2009 to correspond with the year of the large carnivore distribution data, as described above. We chose this dataset because these data provide high resolution estimates of per capita income for both developed and developing countries based on nighttime luminosity data<sup>78</sup>. Nighttime luminosity is important because household survey data (e.g., Afrobarometer or DHS) usually do not contain income data since households fear expropriation from the government and/or cannot provide a monetary equivalent of returns from agricultural production. Generally, as income rises, so too does electricity usage and subsequent nighttime light signature per person, in both production activities and consumptive ones<sup>105</sup>, and light has been used as a proxy for income per capita in several previous studies<sup>106–108</sup>. **Estimating cattle prices**. We obtained data on national cattle prices per kilogram (Item Code 945, FAO) from the Food and Agriculture Organization of the United Nations<sup>109</sup>. This dataset reports on cattle prices as collected at the point of initial sale (prices paid at the farm-gate). Because our large carnivore AOH and per capita income data corresponded to the year 2009, we included the average cattle live weight price for the year 2009, or the next closest year available in the FAO database or gray literature (see Supplementary Data File 1 for details). We then determined the price of a single sub-adult cattle calf by multiplying the per kilogram cattle price for a given country by 250 kg, the approximate size of a sub-adult cattle calf, also known as a single tropical livestock unit<sup>46</sup>. This estimated price of a sub-adult cattle calf per country was used for estimating the financial burden of large carnivore predation, described below.

**Mapping cattle distribution**. We mapped the spatial distribution of cattle using the updated Gridded Livestock of the World (GLW  $3^{110}$ ). The GLW 3 represents areas of sub-national livestock densities, including cattle at fine resolution (~10 km at the equator) for the year 2010. We used the dasymetric dataset of cattle, which corresponds to previous GLW datasets and represents different cattle densities per pixel within a census area according to random forest models. We determine cattle to be present in a given pixel if the density was greater than zero. We resampled the data (300 m).

Mapping burden hotspots. We determined the financial burden of losing a single sub-adult cattle calf to large carnivore predation by first dividing estimates of the price of a single sub-adult cattle calf (see Estimating cattle prices) with estimates of per capita income at the sub-national scale (see *Estimating per capita income*). This produced the relative proportion of annual per capita income lost assuming the predation of a single sub-adult cattle calf. We then masked sub-national administrative boundaries, which contained the above information on the proportion of per capita income vulnerable, with the spatial extent of large carnivores (see Mapping large carnivore habitat)-that is, a spatial mosaic of all 18 large carnivore AOH. Next, we masked this layer with data<sup>110</sup> on the spatial distribution of cattle globally (see Mapping cattle distribution). Finally, we intersected the AOH (corrected with the distribution of cattle) with country boundaries to determine the per capita financial burden at the national scale, and we intersected this with individual carnivore AOH layers to determine the potential financial burden within each species' geographic range (Supplementary Data 3).

We report our results of potential economic burden for each country as an average annual per capita income loss across the entire country. We also calculated the proportion of each carnivore species' range overlapping areas that experience different levels of economic burden: 0-5% of per capita income loss (very low-vulnerability area), 5-10% (low vulnerability area), 10-25% (moderate vulnerability area), 25-50% (high vulnerability area), and >50% (extreme vulnerability area).

**Estimating food security impacts**. We also assessed the direct and opportunity costs lost to households from predation. We calculated calories lost from a single cow or bull (note the FAO data provides hectograms per animal for yield) predation through multiplying country-specific meat yield from the FAO 2021 by beed carcass kilojoule value (1351 kJ/100 g) (see: https://www.fao.org/ag/aginfo/themes/en/meat/backgr\_composition.html). We then divided these lost calories by the average estimated daily caloric intake of a young child aged 2–3, an adolescent aged 12–13, and an adult aged 30–60-years-old<sup>111</sup>.

**Estimating economic disparities in national cattle production**. Finally, we assessed the disparities in cattle prices between developing, transition, and developed economies and examined the differences in meat yield per carcass (Supplementary Data 2). We did this because we wanted to ascertain whether cattle keepers in developing and transition economies may be further exposed (comparatively to developed states) to conflict due to a) price plasticity in cattle markets over time (these could be indicative of market or climate-related shocks<sup>46,48</sup>), and b) low productivity in meat production per animal (ie. prices of cattle per kg may be similar between countries however cattle keepers have to produce more cattle per unit area to yield the same price per ton). We used cattle meat yields in kilograms per animal obtained from the FAO agricultural database<sup>109</sup>. Potential disparities in a) the cattle price (both for the 2009 year, and over time using our historic FAO cattle price dataset), and b) meat yield per animal between developing, transitioning, and developed economies were assessed using a one-way analysis of variance ANOVA.

**Reporting summary**. Further information on research design is available in the Nature Portfolio Reporting Summary linked to this article.

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### References

- Andrade, G. S. & Rhodes, J. R. Protected areas and local communities: an inevitable partnership toward successful conservation strategies? *Ecol. Soc.* 17, 14–23 (2012).
- UNHCR. United Nations High Commissioner for Refugees. The Sustainable Development Goals and Addressing Statelessness (2017). https://www. refworld.org/docid/58b6e3364.html [accessed 16 April 2021]
- Ngorima, A., Brown, A., Masunungure, C. & Biggs, D. Local community benefits from elephants: Can willingness to support anti-poaching efforts be strengthened? *Conserv. Sci. Pract.* 2, e303 (2020).
- Ripple, W. J. et al. Status and ecological effects of the world's largest carnivores. *Science* 343, 1241484 (2014).
- 5. O'Bryan, C. J. et al. The contribution of predators and scavengers to human well-being. *Nat. Ecol. Evol.* 2, 229–236 (2018).
- Levi, T. et al. Community ecology and conservation of bear-salmon ecosystems. Front. Ecol. Evol. 8, 433 (2020).
- Raynor, J. L., Grainger, C. A. & Parker, D. P. Wolves make roadways safer, generating large economic returns to predator conservation. *Proc. Natl Acad. Sci. U.S.A.* 118, e2023251118 (2021).
- Tortato, F. R., Izzo, T. J., Hoogesteijn, R. & Peres, C. A. The numbers of the beast: Valuation of jaguar (Panthera onca) tourism and cattle depredation in the Brazilian Pantanal. *Glob. Ecol. Conserv.* 11, 106–114 (2017).
- Jacobsen, K. S. et al. What is a lion worth to local people—quantifying of the costs of living alongside a top predator. *Ecol. Econ.* 198, 107431 (2022).
- Thirgood, S., Woodroffe, R. & Rabinowitz, A. The impact of human-wildlife conflict on human lives and livelihoods. *Conserv. Biol. Ser.* 9, 13 (2005).
- Mackenzie, C. A. & Ahabyona, P. Elephants in the garden: financial and social costs of crop raiding. *Ecol. Econ.* 75, 72–82 (2012).
- Anaya, F. C. & Espírito-Santo, M. M. Protected areas and territorial exclusion of traditional communities. *Ecol. Soc.* 23 (2018).
- Nsukwini, S. & Bob, U. Protected areas, community costs and benefits: a comparative study of selected conservation case studies from northern KwaZulu-Natal, South Africa. *GeoJ. Tour. Geosites* 27, 1377–1391 (2019).
- Heydinger, J. M., Packer, C. & Tsaneb, J. Desert-adapted lions on communal land: surveying the costs incurred by, and perspectives of, communal-area livestock owners in northwest Namibia. *Biol. Conserv.* 236, 496–504 (2019).
- Dickman, A. J., Macdonald, E. A. & Macdonald, D. W. A review of financial instruments to pay for predator conservation and encourage human-carnivore coexistence. *Proc. Natl Acad. Sci. U.S.A.* 108, 13937–13944 (2011).
- Wang, S. W. & Macdonald, D. W. Livestock predation by carnivores in Jigme Singye Wangchuck National Park, Bhutan. *Biol. Conserv.* 129, 558–565 (2006).
- Holmern, T., Nyahongo, J. & Røskaft, E. Livestock loss caused by predators outside the Serengeti National Park, Tanzania. *Biol. Conserv.* 135, 518–526 (2007).
- Thornton, P. K. et al. Locating poor livestock keepers at the global level for research and development targeting. *Land Use Policy* 20, 311–322 (2003).
- McDermott, J. J., Staal, S. J., Freeman, H. A., Herrero, M. & Van de Steeg, J. A. Sustaining intensification of smallholder livestock systems in the tropics. *Livest. Sci.* 130, 95–109 (2010).
- Dyson-Hudson, N. & Dyson-Hudson, R. The structure of East African herds and the future of East African herders. *Dev. Change* 13, 213–238 (1982).
- Herrero, M. et al. Greenhouse gas mitigation potentials in the livestock sector. Nat. Clim. Change 6, 452–461 (2016).
- Kgathi, D. L., Ngwenya, B. N. & Wilk, J. Shocks and rural livelihoods in the Okavango Delta, Botswana. *Dev. South. Afr.* 24, 289–308 (2007).
- Letta, M., Montalbano, P. & Tol, R. S. Temperature shocks, short-term growth and poverty thresholds: evidence from rural Tanzania. World Dev. 112, 13–32 (2018).
- Cottrell, R. S. et al. Food production shocks across land and sea. *Nat. Sustain.* 2, 130–137 (2019).
- Li, J. et al. Role of Tibetan Buddhist monasteries in snow leopard conservation. *Conserv. Biol.* 28, 87–94 (2014).
- Bhatia, S., Redpath, S. M., Suryawanshi, K. & Mishra, C. The relationship between religion and attitudes toward large carnivores in northern India? *Hum. Dimens. Wildl.* 22, 30–42 (2017).
- Gebresenbet, F., Baraki, B., Yirga, G., Sillero-Zubiri, C. & Bauer, H. A culture of tolerance: coexisting with large carnivores in the Kafa Highlands, Ethiopia. *Oryx* 52, 751–760 (2018).
- 28. Cardillo, M. et al. Human population density and extinction risk in the world's carnivores. *PLoS Biol.* **2**, e197 (2004).
- Hazzah, L., Mulder, M. B. & Frank, L. Lions and warriors: social factors underlying declining African lion populations and the effect of incentivebased management in Kenya. *Biol. Conserv.* 142, 2428–2437 (2009).
- 30. Plaza, P. I., Martínez-López, E. & Lambertucci, S. A. The perfect threat: pesticides and vultures. *Sci. Total Environ.* **687**, 1207–1218 (2019).

## ARTICLE

- 31. Mateo-Tomás, P. & López-Bao, J. V. Poisoning poached megafauna can boost trade in African vultures. *Biol. Conserv.* **241**, 108389 (2020).
- Tumenta, P. N. et al. Threat of rapid extermination of the lion (Panthera leo leo) in Waza National Park, Northern Cameroon. *Afr. J. Ecol.* 48, 888–894 (2010).
- Braczkowski, A. et al. Detecting early warnings of pressure on an African lion (Panthera leo) population in the Queen Elizabeth Conservation Area, Uganda. *Ecol. Solut. Evid.* 1, e12015 (2020b).
- Ickes, K. Hyper-abundance of Native Wild Pigs (Sus scrofa) in a Lowland Dipterocarp Rain Forest of Peninsular Malaysia 1. *Biotropica* 33, 682–690 (2001).
- Ripple, W. J. et al. Widespread mesopredator effects after wolf extirpation. Biol. Conserv. 160, 70–79 (2013).
- Ripple, W. J. & Beschta, R. L. Linking a cougar decline, trophic cascade, and catastrophic regime shift in Zion National Park. *Biol. Conserv.* 133, 397–408 (2006).
- Ripple, W. J. & Beschta, R. L. Trophic cascades involving cougar, mule deer, and black oaks in Yosemite National Park. *Biol. Conserv.* 141, 1249–1256 (2008).
- Gilbert, S. L. et al. Socioeconomic benefits of large carnivore recolonization through reduced wildlife-vehicle collisions. *Conserv. Lett.* 10, 431–439 (2017).
- ILRI. Rangelands Atlas. (ILRI, IUCN, FAO, WWF, UNEP and ILC, 2021). Nairobi Kenva: ILRI.
- 40. Williams, D. R. et al. Proactive conservation to prevent habitat losses to agricultural expansion. *Nat. Sustain.* **4**, 314–322 (2021).
- McManus, J. S. et al. Dead or alive? Comparing costs and benefits of lethal and non-lethal human-wildlife conflict mitigation on livestock farms. *Oryx* 49, 687–695 (2015).
- 42. Broekhuis, F. et al. Identification of human-carnivore conflict hotspots to prioritize mitigation efforts. *Ecol. Evol.* 7, 10630–10639 (2017).
- Lozano, J. et al. Human-carnivore relations: a systematic review. Biol. Conserv. 237, 480–492 (2019).
- Khorozyan, I. & Waltert, M. A global view on evidence-based effectiveness of interventions used to protect livestock from wild cats. *Conserv. Sci. Pract.* 3, e317 (2021).
- Di Minin, E., Slotow, R., Fink, C., Bauer, H. & Packer, C. A pan-African spatial assessment of human conflicts with lions and elephants. *Nat. Commun.* 12, 1–10 (2021).
- Lybbert, T. J. et al. Stochastic wealth dynamics and risk management among a poor population. *Econ. J.* 114, 750–777 (2004).
- Otte, M. J. & Chilonda, P. Cattle and Small Ruminant Production Systems in Sub-Saharan. Africa - Systematic Rev. (FAO, Rome, Italy, 2002).
- Maystadt, J. F. & Ecker, O. Extreme weather and civil war: Does drought fuel conflict in Somalia through livestock price shocks? *Am. J. Agric. Econ.* 96, 1157–1182 (2014).
- Galvin, K. A. Transitions: pastoralists living with change. Annu. Rev. Anthropol. 38, 185–198 (2009).
- 50. Stavi, I. et al. Food security among dryland pastoralists and agropastoralists: The climate, land-use change, and population dynamics nexus. *Anthropocene Rev.* (2021). 20530196211007512.
- Ogra, M. V. Human-wildlife conflict and gender in protected area borderlands: a case study of costs, perceptions, and vulnerabilities from Uttarakhand (Uttaranchal), India. *Geoforum* 39, 1408–1422 (2008).
- Botreau, H., & Cohen, M. J. Gender Inequalities and Food Insecurity: Ten Years After The Food Price Crisis, Why Are Women Farmers Still Food-Insecure? Oxfam:Oxford, UK (2019).
- 53. Salerno, J. et al. Wildlife impacts and changing climate pose compounding threats to human food security. *Curr. Biol.* **31**, 5077–5085 (2021).
- Prado, E. L. & Dewey, K. G. Nutrition and brain development in early life. Nutr. Rev. 72, 267–284 (2014).
- Madhusudan, M. D. The global village: linkages between international coffee markets and grazing by livestock in a south Indian wildlife reserve. *Conserv. Biol.* 19, 411–420 (2005).
- Margulies, J. D. & Karanth, K. K. The production of human-wildlife conflict: A political animal geography of encounter. *Geoforum* 95, 153–164 (2018).
- Simoons, F. J., Simoons, F. I. & Lodrick, D. O. Background to understanding the cattle situation of India: The sacred cow concept in Hindu religion and folk culture. *Zeitschrift Für Ethnologie* 106, 121–137 (1981).
- Good, C., Burnham, D. & Macdonald, D. W. A cultural conscience for conservation. *Animals* 7, 52 (2017).
- Courchamp, F. et al. The paradoxical extinction of the most charismatic animals. *PLoS Biol.* 16, e2003997 (2018).
- Bond, J. & Mkutu, K. Exploring the hidden costs of human-wildlife conflict in northern Kenya. Afr. Stud. Rev. 61, 33–54 (2018).
- Di Minin, E., Leader-Williams, N. & Bradshaw, C. J. Banning trophy hunting will exacerbate biodiversity loss. *Trends Ecol. Evol.* 31, 99–102 (2016).
- Dickman, A. et al. Trophy hunting bans imperil biodiversity. Science 365, 874–874 (2019).

- Bruskotter, J. T., Vucetich, J. A., Gilbert, S. L., Carter, N. H. & George, K. A. Tragic trade-offs accompany carnivore coexistence in the modern world. *Conserv. Lett.* 15, e412841 (2022).
- 64. Dempsey, J. et al. Biodiversity targets will not be met without debt and tax justice. *Nat. Ecol. Evol.* **6**, 237–239 (2022).
- 65. Hallegatte, S. & Rozenberg, J. Climate change through a poverty lens. *Nat. Clim. Change* 7, 250–256 (2017).
- Islam, S. N., and Winkel, J. Climate change and social inequality. DESA Working Paper No. 152. New York, NY: United Nations Department of Economic & Social Affairs (2017).
- Platteau, J. P. Monitoring elite capture in community-driven development. Dev. Change 35, 223-246 (2004).
- Karanth, K. K. & DeFries, R. Nature-based tourism in Indian protected areas: new challenges for park management. *Conserv. Lett.* 4, 137–149 (2011).
- Ament, J. M., Collen, B., Carbone, C., Mace, G. M. & Freeman, R. Compatibility between agendas for improving human development and wildlife conservation outside protected areas: insights from 20 years of data. *People Nat.* 1, 305–316 (2019).
- Geldmann, J., Manica, A., Burgess, N. D., Coad, L. & Balmford, A. A globallevel assessment of the effectiveness of protected areas at resisting anthropogenic pressures. *Proc. Natl Acad. Sci. U.S.A.* 116, 23209–23215 (2019).
- Naidoo, R. et al. Evaluating the impacts of protected areas on human wellbeing across the developing world. *Sci. Adv.* 5, eaav3006 (2019).
- Lichtenfeld, L. L., Trout, C. & Kisimir, E. L. Evidence-based conservation: predator-proof bomas protect livestock and lions. *Biodivers. Conserv.* 24, 483–491 (2015).
- 73. Persson, J., Rauset, G. R. & Chapron, G. Paying for an endangered predator leads to population recovery. *Conserv. Lett.* **8**, 345–350 (2015).
- Barichievy, C. et al. A demographic model to support an impact financing mechanism for black rhino metapopulations. *Biol. Conserv.* 257, 109073 (2021).
- Maingi, S. W. Safari tourism and its role in sustainable poverty eradication in East Africa: the case of Kenya. Worldwide Hosp. Tour. Themes 13, 81–94 (2021).
- Homewood, K. M., Trench, P. C. & Brockington, D. Pastoralist livelihoods and wildlife revenues in East Africa: a case for coexistence? *Pastoralism: Res. Pol. Pract.* 2, 1–23 (2012).
- Thornton, P., Nelson, G., Mayberry, D. & Herrero, M. Impacts of heat stress on global cattle production during the 21st century: a modelling study. *Lancet Planet. Health* 6, e192–e201 (2022).
- Lessmann, C. & Seidel, A. Regional inequality, convergence, and its determinants-a view from outer space. *Eur. Econ. Rev.* 92, 110–132 (2017).
- Brooks, T. M. et al. Measuring terrestrial area of habitat (AOH) and its utility for the IUCN Red List. *Trends Ecol. Evol.* 34, 977–986 (2019).
- Miller, J. R. Mapping attack hotspots to mitigate human-carnivore conflict: approaches and applications of spatial predation risk modeling. *Biodivers. Conserv.* 24, 2887–2911 (2015).
- Gastineau, A., Robert, A., Sarrazin, F., Mihoub, J. B. & Quenette, P. Y. Spatiotemporal depredation hotspots of brown bears, Ursus arctos, on livestock in the Pyrenees, France. *Biol. Conserv.* 238, 108210 (2019).
- 82. Kruuk, H. Surplus killing by carnivores. J. Zool. 166, 233–244 (1972).
- Khorozyan, I. et al. Effects of shepherds and dogs on livestock depredation by leopards (Panthera pardus) in north-eastern Iran. *PeerJ* 5, e3049 (2017).
- Lucherini, M., Guerisoli, M. D. L. M. & Luengos Vidal, E. M. Surplus killing by pumas Puma concolor: rumours and facts. *Mammal. Rev.* 48, 277–283 (2018).
- Ocaido, M., Muwazi, R. T. & Opuda-Asibo, J. Financial analysis of livestock production systems around Lake Mburo National Park, in South Western Uganda. *Livest. Res. Rural Dev.* 21, 70 (2009).
- Dyson-Hudson, R. & Dyson-Hudson, N. Nomadic pastoralism. Annu. Rev. Anthropol. 9, 15–61 (1980).
- Barber, J. P. The Karamoja District of Uganda: a pastoral people under colonial rule. J. Afr. Hist. 3, 111–124 (1962).
- Oberg, K. Analysis of the Bahima marriage ceremony. Africa 19, 107–120 (1949).
- Purseglove, J. W. Banyankole Agriculture. East Afr. Agric. J. 5, 198–207 (1939).
- 90. Canonici, N. N. Food in Zulu folktales. South. Afr. J. Folk. Stud. 2, 24–36 (1991).
- United Nations. World Population Prospects 2022: Summary of Results. UN DESA/POP/2022/TR/NO. 3 (2022).
- Tomas, W. M. et al. Sustainability agenda for the Pantanal Wetland: perspectives on a collaborative interface for science, policy, and decisionmaking. *Trop. Conserv. Sci.* 12, 1940082919872634 (2019).
- Vale, P. et al. Mapping the cattle industry in Brazil's most dynamic cattleranching state: Slaughterhouses in Mato Grosso, 1967-2016. PLOS One 14, e0215286 (2019).

- 94. De Leeuw, P. N., Bekure, S., & Grandin, B. E. Aspects of livestock productivity in Maasai group ranches in Kenya. ILCA Bull. (1984).
- Barua, M., Bhagwat, S. A. & Jadhav, S. The hidden dimensions of human-wildlife conflict: health impacts, opportunity and transaction costs. *Biol. Conserv.* 157, 309–316 (2013).
- Choudhury, A. Human–elephant conflicts in Northeast India. *Hum. Dimens.* Wildl. 9, 261–270 (2004).
- 97. Sherman, P. B., & Dixon, J. A. Economics of protected areas: a new look at benefits and costs. *Earthscan Publications Limited* (1990).
- Braczkowski, A. et al. Evidence for increasing human-wildlife conflict despite a financial compensation scheme on the edge of a Ugandan National Park. *Conserv. Sci. Pract.* 2, e309 (2020c).
- Gulati, S., Karanth, K., Nguyet Anh Le, N. & Noack, F. Human casualties are the dominant cost of human-wildlife conflict in India. *Proc. Natl Acad. Sci.* 118, e1921338118 (2021).
- Rondinini, C. et al. Global habitat suitability models of terrestrial mammals. *Philos. Trans. R. Soc.* 366, 2633–2641 (2011).
- 101. Lewis, J. S. et al. Biotic and abiotic factors predicting the global distribution and population density of an invasive large mammal. *Sci. Rep.* 7, 44152 (2017).
- 102. Strassburg, B. et al. Global priority areas for ecosystem restoration. *Nature* 586, 724–729 (2020).
- 103. O'Bryan, C. J. et al. The importance of indigenous peoples' lands for the conservation of terrestrial mammals. *Conserv. Biol.* **35**, 1002–1008 (2021).
- 104. Rondinini, C., Wilson, K. A., Boitani, L., Grantham, H. & Possingham, H. P. Tradeoffs of different types of species occurrence data for use in systematic conservation planning. *Ecol. Lett.* 9, 1136–1145 (2006).
- Henderson, J. V., Storeygard, A. & Weil, D. N. Measuring economic growth from outer space. Am. Econ. Rev. 102, 994–1028 (2012).
- 106. Ebener, S., Murray, C., Tandon, A. & Elvidge, C. C. From wealth to health: modelling the distribution of income per capita at the sub-national level using night-time light imagery. *Int. J. Health Geographics* 4, 5 (2005).
- 107. Chen, X. & Nordhaus, W. D. Using luminosity data as a proxy for economic statistics. *Proc. Natl Acad. Sci.* 108, 8589–8594 (2011).
- 108. Jean, N. et al. Combining satellite imagery and machine learning to predict poverty. *Science* 353, 790–794 (2016).
- 109. FAO Meat live weight, cattle database. License: CC BY-NC-SA 3.0 IGO. http:// www.fao.org/faostat/en/#search/cattle (2021). Accessed 24 April 2021.
- 110. Gilbert, M. et al. Global distribution data for cattle, buffaloes, horses, sheep, goats, pigs, chickens and ducks in 2010. *Sci. Data* **5**, 1–11 (2018).
- 111. United Nations University & World Health Organization. Human Energy Requirements: Report of a Joint FAO/WHO/UNU Expert Consultation: Rome, 17–24 October 2001 (Vol. 1) Food & Agriculture Org (2004).

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#### **Competing interests**

The authors declare no competing interests.

#### Additional information

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